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# Determination of Reaction Rate constant for p-Chlorophenol and Nitrobenzene reacting with <sup>•</sup>OH during oxidation by Fe(II)/H<sub>2</sub>O<sub>2</sub> system

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**Abstract:** Fenton's oxidation process (using a mixture of hydrogen peroxide and ferrous ion) rapidly oxidizes these organic compounds under acidic conditions. Different molar ratios of  $H_2O_2$  to the organic compound were used. This study describes oxidation of 4-chlorophenol and nitrobenzene, in aqueous solutions, as individual pollutant. A ratio of  $[H_2O_2] / [Fe^{2+}]$  of 8.2 was kept constant for all the experiments based on typical ratios reported in literature. Maximum 99.29% degradation for p-chlorophenol and 96.32% degradation for nitrobenzene were achieved at  $[H_2O_2] / [organic]$  molar ratio of 7.55 and 3.60 respectively. In this work, the experimental initial rate data is fitted to a kinetic model for determination of rate constant values between organic compound and hydroxyl radicals for individual compounds. The second order absolute rate constant for reaction between hydroxyl radical and p-chlorophenol is found to be 2.144 ×10<sup>6</sup>/M-s, and the same in case of nitrobenzene is:  $1.102 \times 10^6/M$ -s. **Keywords:** Kinetic model; Initial degradation rate; Rate constant estimation, Fenton's reaction.

# **1. INTRODUCTION**

Chlorophenols (CP) are introduced into the environment due to several man-made activities, such as water disinfection, waste incineration and also as byproducts in the bleaching of paper pulp with chlorine <sup>1</sup> and have become one of the major contaminants in water and air due to their non-biodegradability.<sup>2,3</sup> and carcinogenicity All chlorophenols possess bactericidal activities, phytotoxicity and have ability to bioaccumulate in organisms and these properties tend to increase with increasing chlorination.<sup>4</sup> Parachlorophenol / 4chlorophenol is a solvent for extraction of sulfur and nitrogen from coal, as an intermediate in the synthesis of dyes and drugs or solvent in the refining of oils. It is found in industrial wastewaters from oil refineries,

iron smelting, resin, manufacturing etc and it is also listed among the 65 priority pollutants by the USEPA.<sup>5</sup> p-chlorophenol was therefore chosen as a model compound in this study because of its high toxicity and being largely non-biodegradable, posing serious risks to the environment.<sup>6</sup>

Nitrobenzene, an organic compound, consisting of a benzene ring and a 'nitro' functional group has been categorized as hazardous waste by USEPA.<sup>7</sup> It is widely used in the production of products, such as dyes,<sup>8</sup> explosives and pesticides <sup>9</sup> and for preparation of textiles and papers. Consequently, this compound is a common water pollutant due to its presence in industrial wastewater.<sup>10,11</sup> It affects the central nervous systems, causes fatigue, headache, vomiting and coma.

In this work, nitrobenzene (NB), which is also listed as a priority pollutant by USEPA was also chosen as a model pollutant.

The conventional technologies for the removal of both p-chlorophenol and nitrobenzene include biological treatment, incineration, adsorption over activated etc.<sup>12</sup> stripping, Remediation carbon, air of wastewaters containing these pollutants is very difficult by biological treatment because activities of microorganisms are affected by toxicity; incinerations may cause substantial emission of other hazardous compounds; adsorption or air stripping requires a posttreatment to remove the pollutants from the newly contaminated environment. Although these methods are effective, there are some difficulties in using these processes: high cost, formation of hazardous byproducts and applicability to only a limited concentration range. 13-15

Among the various methods available, the advanced oxidation processes (AOPs) appear to be a promising technology which have been reported to be effective at ambient condition in the degradation of wide variety of soluble organic contaminants from wastewaters.<sup>16-19</sup>

Fenton's oxidation, one of the AOPs that involve the generation of the highly reactive hydroxyl radical ( $^{\circ}$ OH) having stronger oxidation potential (2.80 V) than ozone (2.07 V), is of current interest for the destruction of of organic pollutants in synthetic wastewaters. Fenton's reagent (hydrogen peroxide in presence of ferrous salt) involves homogenous reaction and is environmentally acceptable. The Fenton's reactions at acidic pH lead to the production of ferric ion (Fe<sup>3+</sup>) and of the hydroxyl radicals ( $^{\circ}$ OH).<sup>20</sup>

The thrust in study regarding oxidation of organic compounds by Fenton's reagent has mostly been on studying the effects of concentration (substrate and oxidant) and catalyst (FeSO<sub>4</sub>) on the degradation rates. In this present work, a kinetic model from fundamentals is fitted. Reaction rate constants between individual organic compounds (B) and <sup>•</sup>OH are determined by fitting the model to the experimental data. Proper kinetic data would enable modeling, design and optimization of reactors for these systems. The main objective of the present work is to study the of degradation kinetics p-chlorophenol and nitrobenzene by Fenton's reaction, when they are present separately and evaluate the hitherto unreported

kinetic constants for degradation of the said organics by hydroxyl radicals.

# 2. MATERIALS AND METHODS

# **2.1. REAGENTS**

P-chlorophenol and nitrobenzene (both GR grade), Hydrogen peroxide solution (50%, v/v), Ferrous sulfate heptahydrate (FeSO<sub>4</sub>.7H<sub>2</sub>O) were procured from M/s Merck, India. NaOH and H<sub>2</sub>SO<sub>4</sub> were used to adjust the pH of the solution. Distilled water was used for preparing aqueous solutions throughout.

#### **2.2. EXPERIMENTAL PROCEDURE**

All experiments were carried out in batch mode in beakers of 1000 ml capacity with 500 ml of pollutant solution. Appropriate amount of FeSO<sub>4</sub> was dissolved using part of the liquid in the beaker and then the same was transferred back to beaker. To start each experiment, appropriate amount of  $H_2O_2$  were added into the beaker being stirred with a magnetic stirrer at a fixed high (250) rpm. Initial pH of the reaction solution was adjusted by using 1M  $H_2SO_4$ , as and when necessary. Small volumes of samples were withdrawn at interval of 5 minutes using a pipette. 0.1 ml of 4 N NaOH per ml of sample was added to arrest the reaction and then the sample was filtered through Whatman No. 42 filter paper. The filtrate was analyzed by gas chromatography.

# **2.3 ANALYTICAL METHODS**

pH was recorded using a digital pH meter (Model: CL-46, Toshniwal Instruments, India). P-chlorophenol and nitrobenzene concentrations were determined using a gas chromatograph (Chemito India Ltd. GC 8610 with a Chemito 10% OV17 column, 8' x 0.125") with Flame Ionization Detector (FID). Carrier gas (Nitrogen) flow was kept at 25 ml/min and the sample volume being injected was 0.5  $\mu$ l. Other operating conditions were: Injector temperature 230°C, Column temperature 120-220 °C and Detector temperature 280°C. Pchlorophenol and nitrobenzene peaks were observed at a retention time of 3.8 minutes and 3.2 minutes.

#### 3. RESULTS AND DISCUSSION

The results of two sets of experiments are presented in this section.

Exp.	$[B_1] \ge 10^{-3}$	$[H_2O_2]$	$[Fe^{2+}] \times 10^{-4}$	$R = [H_2O_2] / [B_1]$	$-(d[B_1]/dt)_0 \ge 10^4$
No.	gmol/l	gmol/l	gmol/l	mol/mol	(gmol/l-s)
E-1	3.889	0.0220	26.78	5.65	0.354
E-2	5.445	0.0308	37.50	5.65	0.559
E-3	7.778	0.044	53.57	5.65	1.04
E-4	3.889	0.0294	35.71	7.55	0.227
E-5	5.445	0.04116	50	7.55	0.698
E-6	7.778	0.0588	71.42	7.55	1.390

Table 1: Initial conditions for the degradation of p-chlorophenol (B<sub>1</sub>) for [H<sub>2</sub>O<sub>2</sub>]/[Fe<sup>2+</sup>] molar ratio 8.2

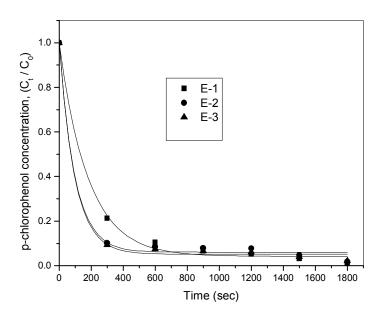


Figure 1. Fall in p-Chlorophenol concentration with time for different initial organic concentrations and same ratio (5.65) of hydrogen peroxide to p-Chlorophenol

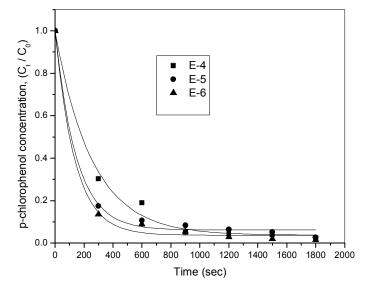


Figure 2. Fall in p-Chlorophenol concentration with time for different initial organic concentrations and same ratio (7.55) of hydrogen peroxide to p-Chlorophenol

Time	E-1	E-2	E-3	E-4	E-5	E-6	
(min)	pН	pН	pH	pН	pН	pН	
0	3.12	3.17	3.10	3.09	2.98	2.97	
5	2.92	2.81	2.75	2.80	2.76	2.62	
10	2.77	2.75	2.63	2.73	2.63	2.54	
15	2.56	2.71	2.61	2.68	2.60	2.52	
20	2.44	2.69	2.57	2.67	2.58	2.50	
25	2.41	2.64	2.56	2.66	2.57	2.49	
30	2.38	2.62	2.55	2.66	2.57	2.48	

Table 2: pH of solution with time during p-chlorophenol degradation

Table 3: Initial conditions for the degradation of nitrobenzene (B<sub>2</sub>) for [H<sub>2</sub>O<sub>2</sub>] / [Fe<sup>2+</sup>] molar ratio 8.2

Exp.No.	[B <sub>2</sub> ] x 10 <sup>-3</sup>	[H <sub>2</sub> O <sub>2</sub> ]	$[Fe^{2+}] \ge 10^{-4}$	$[H_2O_2]/[B_2] = R$	$-(d[B_2]/dt)_0 \ge 10^5$
	gmol/l	gmol/l	gmol/l	mol/mol	(gmol/l-s)
E-7	3.25	0.0117	14.28	3.6	0.731
E-8	4.87	0.0176	21.42	3.6	1.26
E-9	3.25	0.0235	28.57	7.2	1.05
E-10	4.06	0.044	53.57	10.8	2.25
E-11	4.87	0.0529	64.28	10.8	4.63

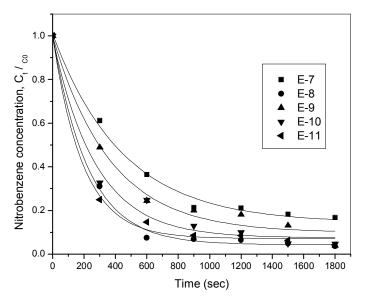


Figure 3. Fall in Nitrobenzene concentration with time for different initial organic concentrations

I able 4: pH of solution with time during nitrobenzene degradation							
Time(min)	E-7 pH	E-8 pH	E-9 pH	E-10 pH	E-11 pH		
0	3.11	3.02	3.02	3.06	3.10		
5	3.07	3.0	3.02	3.02	3.05		
10	3.07	2.96	2.98	3.02	3.0		
15	3.02	2.95	2.98	3.0	3.0		
20	3.01	2.94	2.98	2.99	3.0		
25	2.96	2.94	2.98	2.95	2.95		
30	2.96	2.93	2.97	2.95	2.94		

Table 4:	nH	of solution	with	time	during	nitrobenzene	degradation
1 and <b>7</b> .	PTT	or solution	****	umu	uuiing	IIIII UDUILLIIL	ucziauanon

#### **<u>3. RESULTS AND DISCUSSION</u>**

The results of two sets of experiments are presented in this section.

#### **3.1. DEGRADATION OF AQUEOUS p-CHLOROPHENOL**

Six experiments were conducted for degradation of aqueous p-chlorophenol by varying the (1) initial concentration of organic compound (3.889 x  $10^{-3}$  gmol/l to 7.778 x  $10^{-3}$  gmol/l), (2) molar ratio of H<sub>2</sub>O<sub>2</sub> to organic component (5.65 to 7.55). The molar ratio of H<sub>2</sub>O<sub>2</sub> to Fe<sup>2+</sup> was kept constant at 8.2, based on typical range of this parameter reported in literature.<sup>21</sup> Initial conditions of all six experiments are summarized in Table 1. It also reports the rate of degradation of the organic compound determined from initial slope of its concentration-time profile.

Fig. 1 & 2 shows the fall in p-chlorophenol concentration (expressed as a fraction of initial concentration) with time. The change in pH of solution with time is presented in Table 2.

#### **3.2. DEGRADATION OF NITROBENZENE**

Five experiments were conducted for degradation of aqueous nitrobenzene by varying the (1) initial concentration of organic species ( $3.25 \times 10^{-3}$  gmol/l to  $4.87 \times 10^{-3}$  gmol/l), (2) molar ratio of H<sub>2</sub>O<sub>2</sub> to organic component (3.6 to 10.8). Molar ratios of H<sub>2</sub>O<sub>2</sub> to Fe<sup>2+</sup> in these experiments were kept constant. Initial conditions of all five experiments are presented in Table 3 along with the rate of degradation of the organic compound determined from initial slope of its concentration-time profile.

Fig. 3 shows the fall in concentration of nitrobenzene (expressed as a fraction of initial concentration), with time and the change in pH of solution with time is presented in Table 4.

#### **3.3. KINETIC MODELING**

The oxidation mechanism of organic compound in a single step mineralization by Fenton's reagent is thought to occur in following stages.<sup>22</sup>

$$B + H_2 O_2 \xrightarrow{K_2} B_{OX} \dots$$
 .....(2)

$$H_2O_2 + OH^{\bullet} \xrightarrow{K_4} HO_2^{\bullet} + H_2O \qquad \dots \dots \dots (4)$$

$$OH^{\bullet} + Fe^{2+} \xrightarrow{K_5} Fe^{3+} + OH^{-} \qquad \dots \dots (5)$$

$$Fe^{3+} + H_2O_2 \xrightarrow{\kappa_6} Fe^{2+} + H^+ + HO_2^{\bullet} \quad \dots \dots (6)$$

# **3.4. RATE CONSTANT DETERMINATION**

As it is observed, 'OH are generated or consumed in equations (1), (3), (4) and (5). Therefore, the net rate of generation can be expressed as follows:

$$\frac{d[OH]}{dt} = k_1[H_2O_2][Fe^{2+}] - k_3[B][OH] - k_4[H_2O_2][OH] - k_5[OH][Fe^{2+}]$$
.....(8)

Considering the concentration of the highly reactive 'OH radical remains stationary, the left hand side of Equation (7) is 'zero'. Rearrangement of the terms of Equation (8) gives the concentration of hydroxyl radicals.

['OH] = 
$$\frac{k_1[H_2O_2][Fe^{2^+}]}{k_3[B] + k_4[H_2O_2] + k_5[Fe^{2^+}]} \dots$$

As the organic compound is degraded by reactions (2) and (3), the rate of degradation of organic compound (B) may be expressed as

On substitution the expression for 'OH radicals in equation (9), it becomes

$$-\frac{d[B]}{dt} = k_2[B] [H_2O_2] + k_3 [B] \frac{k_1[H_2O_2][Fe^{2+}]}{k_3[B] + k_4[H_2O_2] + k_5[Fe^{2+}]}$$
.....(11)

This is the rate expression for the oxidation of organic compound which includes five reaction rate constants. Experiments with only  $H_2O_2$  were conducted. Practically no degradation of the organic species was observed in presence of  $H_2O_2$  alone. So  $k_2$  was considered negligible in this model. Among the other rate constants, the literature <sup>23-25</sup> reported value of rate constant  $k_1$  is 76 /M-s,  $k_4$  is (3 x 10<sup>7</sup> /M-s) and the  $k_5$  value is (3 x 10<sup>8</sup> /M-s). Our present study aims at evaluation of the rate constants ( $k_3$ ) for the organic compounds and hydroxyl radicals.

Eq. (11) is rewritten in the form

$$\left(-\frac{d[B]}{dt}\right)_{0} = \frac{k_{1}k_{3} [B]_{0} [H_{2}O_{2}]_{0} [Fe^{2^{+}}]_{0}}{k_{3}[B]_{0} + k_{4} R [B]_{0} + k_{5} [Fe^{2^{+}}]_{0}} \dots \dots (12)$$

Where R is the ratio  $[H_2O_2]_0 / [B]_0$ 

The rate constant  $k_3$  for the two compounds were determined by minimizing the sum of square of deviation of the experimental initial rate and the model

.....(9)

predicted rate for all the experiments in Table 1 for pchlorophenol and in table 3 for nitrobenzene respectively. shows the experimental and model predicted initial degradation rate for p-chlorophenol and nitrobenzene respectively. Average % deviation is below +/-15% as evidenced by the error limit lines of Fig. 4 & 5.

MATLAB routine 'LSQNONLIN' was used to solve this non-linear least square fitting problem. Fig. 4 & 5

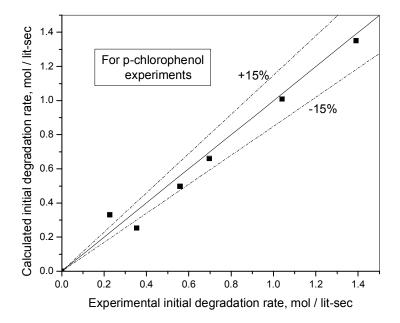


Figure 4. Experimental and model predicted initial degradation rate for p- chlorophenol

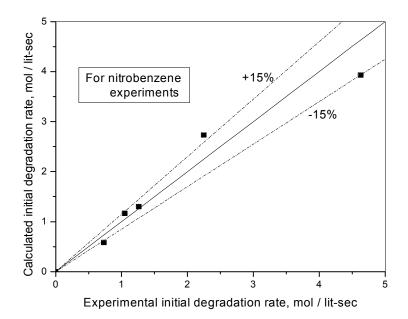


Figure 5. Experimental and model predicted initial degradation rate for nitrobenzene

#### **4. CONCLUSION**

The results obtained in this work demonstrate that aqueous solutions containing p-chlorophenol or nitrobenzene can be effectively degraded by Fenton oxidation process. The reasonably high values of rate constants for reaction of 'OH with p-chlorophenol  $(2.144 \times 10^6 / \text{M-s})$  and with nitrobenzene  $(1.102 \times 10^6 / \text{M-s})$  suggest Fenton's reaction to be a fast and efficient method for degrading these organic pollutants in low concentration. Experimental results show

### 5. REFERENCES

- 1. Ahlborg U.G. and Thunberg T.M., Chlorinated phenols: occurrence, toxicity, metabolism and environmental impact, Crit. Rev. Toxicol., 1980, 7, 1-35.
- 2. Gimeno O., Carbajo M., Beltran F.J. and Rivas F.J., Phenol and substituted phenols AOPs remediation, J. Hazard. Mater., 2005, 119, 99-108.
- 3. Pi Y., Zhang L. and Wang J., The formation and influence of hydrogen peroxide during ozonation of para-chlorophenol, J. Hazard. Mater., 2007, 141, 707-712.
- 4. Titus M.P., Molina V.G., Banos M.A., Gimenez J. and Esplugas S., Degradation of chlorophenols by means of advanced oxidation processes: a general review, Appl. Catal. B: Environ., 2004, 47, 219-256.
- Keith L.H. and Telliard W.A., Priority pollutants: I. A prospective view, Environ. Sci. Technol., 1979, 13, 416-423.
- 6. Hugul M., Boz I. and Apak R., Photocatalytic decomposition of 4-chlorophenol over oxide catalysts, J. Hazard. Mater., 1999, 64, 313-322.
- 7. Sittig M., Handbook of Toxic and Hazardous Chemicals and Carcinogens. 2nd ed, NP Noyes Publications, Park Ridge, NJ, US, 1985, 662-664.
- Contreras S., Rodriguez M., Chamarro E. and Esplugas S., UV- and UV/Fe (III)-enhanced ozonation of nitrobenzene in aqueous solution, J. Photochem. Photobiol. A: Chem., 2001, 142, 79-83.
- Rodriguez M., Kirchner A., Contreras S., Chamarro E. and Esplugas S., Influence of H<sub>2</sub>O<sub>2</sub> and Fe (III) in the photodegradation of nitrobenzene, J. Photochem. Photobiol. A: Chem., 2000, 133, 123-127.

reduction of p-chlorophenol concentration from 7.77 x  $10^{-3}$  gmol/l to 1.232 x  $10^{-3}$  gmol/l and concentration of nitrobenzene falling from 4.87 x  $10^{-3}$  gmol/l to 0.179 x  $10^{-4}$  gmol/l within half an hour. Good fit of the experimental results and the theoretical model justifies the assumption that degradation of organics by H<sub>2</sub>O<sub>2</sub> is negligible compared to the same by 'OH during Fenton's process. The kinetic study indicated that the degradation of these compounds followed the second order kinetics.

- Beltran F.J., Encinar J.M. and Alonso M.A., Nitroaromatic hydrocarbon ozonation in water. 1. Single ozonation, Ind. Eng. Chem. Res., 1998, 37, 25-31.
- 11. Sarasa J., Roche M.P., Ormad M.P., Gimeno E., Puig A. and Ovelleiro J.L., Treatment of a wastewater resulting from dye manufacturing with ozone and chemical coagulation, Water Res., 998, 32, 2721-2727.
- 12. Jardin W.F., Moraes S.G. and Takiyama M.M.K., Photocatalytic degradation of aromatic chlorinated compounds using TiO2: Toxicity of intermediates, Water Res., 1997, 31, 1728-1732.
- 13. Milstein O., Removal of chlorophenols and chlorolignins from bleaching effluent by combined chemical and biological treatment, Water Sci. Technol., 1988, 20, 161-170.
- 14. Bryant W., The removal of chlorinated organic from conventional pulp and paper wastewater treatment systems, Water Sci. Technol., 1992, 26, 417-425.
- 15. Wang R.C., Kuo C.C. and Shyu C.C., Adsorption of phenols onto granular activated carbon in a liquid-solid fluidized bed, J. Chem. Technol. Biotechnol., 1997, 68, 187-194.
- 16. Guittonneau S., De Laat J., Dore M., Duguet J.P. and Bonnel C., Comparative study of the photodegradation of aromatic compounds in water by UV and H2O2/UV, Environ. Technol. Lett., 1988, 9, 1115-1128.
- 17. Legrini O., Oliveros E. and Braun A.M., Photochemical processes for water treatment. Chem. Rev., 1993, 93, 671-698.
- Masten S.J. and Davies S.H.R., The use of ozonation to degrade organic contaminants in wastewaters, Environ. Sci. Technol., 1994, 28, 180-185.

- Andreozzi R., Caprio V., Insola A. and Marotta R., Advanced oxidation processes (AOP) for water purification and recovery, Catal. Today, 1999, 53, 51-59.
- 20. Gallard H., De Laat J. and Legube B, Spectrophotometric study of the formation of iron(III)- hydroperoxy complexes in homogeneous aqueous solution, Water Res., 1999, 33, 2929-2936.
- 21. Ahmadi M., Vahabzadeh F., Bonakdarpour B., Mofarrah E. and Mehranian M, Application of the central composite design and response surface methodology to the advanced treatment of olive oil processing wastewater using Fenton's peroxidation, J. Hazard. Mater., 2005, B123, 187-195.
- 22. De Heredia J. B., Torregrosa J., Dominguez J.R. and Peres J.A., Kinetic model for phenolic compound oxidation by Fenton's reagent, Chemosphere, 2001, 45, 85-90.
- Suna J.H., Suna S.P., Fan M.H., Guoa H.Q., Qiao L.P. and Suna R.X., A kinetic study on the degradation of p-nitroaniline by Fenton oxidation process, J. Hazard. Mater., 2007, 148, 172-177.
- 24. Duesterberg C.K., Cooper W.J. and Waite T.D., Fenton-Mediated Oxidation in the Presence and Absence of Oxygen, Environ. Sci. Technol., 2005, 39, 5052-5058.
- 25. Kwan W.P. and Voelker B.M., Decomposition of Hydrogen Peroxide and Organic Compounds in the Presence of Dissolved Iron and Ferrihydrite, Environ. Sci. Technol., 2002, 36, 1467-1476.

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